



# Effects of NO<sub>2</sub> and RH on secondary organic aerosol formation and light absorption from OH oxidation of *o*-xylene

Shijie Liu<sup>a,b</sup>, Yiqian Wang<sup>a</sup>, Xinbei Xu<sup>a</sup>, Gehui Wang<sup>a,b,\*</sup>

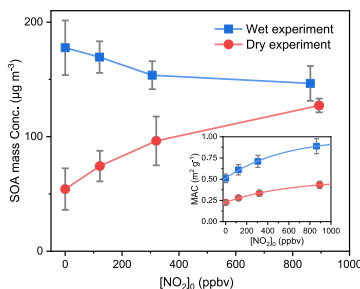
<sup>a</sup> Key Lab of Geographic Information Science of the Ministry of Education, School of Geographic Sciences, East China Normal University, Shanghai, 210062, China

<sup>b</sup> Institute of Eco-Chongming, 3663 North Zhongshan Road, Shanghai, 200062, China

## HIGHLIGHTS

- A chamber study of the compound effect of NO<sub>2</sub> and RH on the *o*-xylene SOA formation through OH-oxidation.
- The evolution characteristics of the *o*-xylene SOA chemical compositions were studied by the HR-ToF-AMS.
- The *o*-xylene SOA mass concentration increased with initial NO<sub>2</sub> concentration at dry condition, but decreased at high RH condition.
- Both the increasing NO<sub>2</sub> concentration and RH condition is beneficial to increase the MAC value of *o*-xylene-derived SOA.

## GRAPHICAL ABSTRACT



## ARTICLE INFO

Handling Editor: Volker Matthias

### Keywords:

*o*-xylene  
Secondary organic aerosol  
Nitrogen dioxide  
Mass absorption coefficient  
Aerosol mass spectrometry

## ABSTRACT

*O*-xylene is an important aromatic volatile organic compound (VOC) in the atmosphere over urban areas. In this work, the effect of nitrogen dioxide (NO<sub>2</sub>) concentration and relative humidity (RH) on the mass concentration of secondary organic aerosols (SOA) formed from *o*-xylene OH oxidation was investigated in a photooxidation chamber. The *o*-xylene SOA mass concentration increased from 54.2 μg m<sup>-3</sup> to 127.2 μg m<sup>-3</sup> during dry conditions, but decreased from 177.7 μg m<sup>-3</sup> to 146.5 μg m<sup>-3</sup> during high RH conditions when the initial NO<sub>2</sub> concentration increased from 0 ppbv to about 900 ppbv. An increase in the ratio of [NO<sub>3</sub>]/[Org] and a decrease in the oxidation state of carbon (OS<sub>C</sub>) of SOA suggested that acid-catalyzed heterogeneous reaction was responsible for enhancing SOA formation with increasing NO<sub>2</sub> concentrations in dry conditions. In contrast, in humid conditions, the high molecular diffusion capacity of SOA could promote the reactivity of OH towards the interior of SOA, and the enhancement of nitrous acid (HONO) formation under high NO<sub>2</sub> conditions could promote the SOA aging processes and be responsible for the decreasing trend of SOA formation with NO<sub>2</sub>. Light absorption by SOA was also measured, and both NO<sub>2</sub> and RH enhanced the mass absorption coefficient (MAC<sub>λ = 365 nm</sub>) value for the optical properties of *o*-xylene SOA. The highest MAC<sub>λ = 365 nm</sub> value of *o*-xylene SOA was 0.89 m<sup>2</sup> g<sup>-1</sup>, observed during humid conditions with an initial NO<sub>2</sub> concentration of 862 ppbv, which was 3.9 times higher than in the experiment conducted in the absence of NO<sub>2</sub> under dry conditions. The formation of nitrogen-containing organic compounds (NOCs) and humic-like substances (HULIS) were responsible for the increased MAC<sub>λ = 365 nm</sub> values of *o*-xylene derived SOA. This study provides new insight into the effect of NO<sub>2</sub> on

\* Corresponding author. Key Lab of Geographic Information Science of the Ministry of Education, School of Geographic Sciences, East China Normal University, Shanghai, 210062, China.

E-mail address: [ghwang@geo.ecnu.edu.cn](mailto:ghwang@geo.ecnu.edu.cn) (G. Wang).

<https://doi.org/10.1016/j.chemosphere.2022.136541>

Received 11 July 2022; Received in revised form 13 September 2022; Accepted 16 September 2022

Available online 20 September 2022

0045-6535/© 2022 Elsevier Ltd. All rights reserved.

SOA formation through the change in *o*-xylene photooxidation under different RH conditions, and the complex effect of multiple environmental factors on SOA formation was also important and should not be ignored.

## 1. Introduction

Secondary organic aerosols (SOA) are an important component of atmospheric aerosols and have adverse effects on ambient air quality and human health (George et al., 2015; Gu et al., 2021). The class of organic aerosol called brown carbon (BrC), which exhibits exponentially increasing absorption in the light spectrum from the visible to ultraviolet (UV) wavelength, could also affect the solar radiation balance and even regional climates (Noziere et al., 2015; Zhang et al., 2017; Liu et al., 2021c). The importance of SOA in the atmospheric environment has been widely recognized. However, due to the diversity of SOA precursors, reaction pathways, and environment impact factors (Liu et al., 2021b; Gaubert et al., 2021; Aruffo et al., 2022), the mechanism of SOA formation are still not clear, and attempts at large-scale atmospheric simulations have not achieved good agreement with SOA concentrations from field observations (Charan et al., 2019; Huang et al., 2020; Gaubert et al., 2021).

Nitrogen oxides ( $\text{NO}_x = \text{NO} + \text{NO}_2$ ) are not only the predominant gaseous pollutant in the atmosphere, but also the most important driver of atmospheric oxidation reactions (Zhang et al., 2015). The photolysis of  $\text{NO}_x$  at different levels has been closely linked to the formation and concentrations of main atmospheric oxidants (such as OH and  $\text{O}_3$ ) in the atmosphere, which participate in the initial oxidation steps of SOA formation (Li et al., 2022). The fate of peroxy radical ( $\text{RO}_2$ ) intermediates produced by photooxidation of volatile organic compounds (VOCs) was also linked to  $\text{NO}_x$  levels (Sarrafzadeh et al., 2016; Zhao et al., 2018). With increasing  $\text{NO}_x$  concentration,  $\text{RO}_2$  react more with NO, rather than with  $\text{HO}_2$ , and are rapidly converted to the alkoxy radicals (RO), which produces more high volatile compounds through fragmentation to suppress SOA formation. However,  $\text{NO}_x$  has also been widely reported to promote SOA formation. In recent years, highly oxygenated organic molecules (HOMs) derived from  $\text{RO}_2$  autooxidation have been considered another key to SOA formation processes. An alkyl radical ( $\text{R}\cdot$ ) with hydroperoxide functionality will form through hydrogen (H)-shift isomerization reactions from  $\text{RO}_2$  in the presence of NO. New  $\text{RO}_2$  will be formed by adding  $\text{O}_2$  to  $\text{R}\cdot$ . HOMs with low-volatility compounds would then be formed through repetition of the autooxidation process and promote SOA formation in the presence of  $\text{NO}_x$  (Bianchi et al., 2019; Li et al., 2022). Nitrogen-containing organic compounds (NOCs) could also be formed through the  $\text{RO}_2 + \text{NO}$  reaction, and the effects of NOCs on the nucleation, concentrations, and physicochemical properties of SOA should not be ignored (Ng et al., 2007a; Xu et al., 2015). Liu et al. (2021c) pointed out that the mass concentration ratio of nitrate ions to SOA increased with increasing initial  $\text{NO}_2$  concentration, which suggested that the acid-catalyzed reaction would be enhanced by an increase in the  $\text{NO}_2$  concentration, which might another important mechanism underlying the promotion effect of  $\text{NO}_2$  on SOA formation (Liu et al., 2021c).

Although the effect of  $\text{NO}_2$  on SOA formation has been widely reported in previous experimental studies, the combined impacts of  $\text{NO}_2$  and other environmental factors on the formation of SOA have not received sufficient attention. Liu et al. (2021b) reported that more semi-volatile oxidation products were formed at higher  $\text{NO}_2$  concentrations, and they inhibited SOA yield. However, the presence of  $\text{NH}_3$  increased the distribution efficiency of semi-volatile oxidation products in the particle phase, resulting in an increased SOA yield. When  $\text{SO}_2$  was present, it was converted to sulfuric acid, which promoted SOA formation by increasing aerosol acidity (Jang et al., 2002; Chu et al., 2016). In the photooxidation process, sulfate formation was enhanced at high  $\text{NO}_2$  concentrations, and some studies have pointed out that this enhancement of SOA yield due to particle acidity was only observed in high- $\text{NO}_2$

conditions rather than in low- $\text{NO}_2$  conditions (Offenberg et al., 2009; Eddingsaas et al., 2012; Wang et al., 2016a; Liu et al., 2019a). The role of water in SOA formation has also been inconsistent among studies. Jia and Xu (2018) pointed out that the oligomerization reaction of stable Criegee intermediates (SCIs) from the  $\text{O}_3$  channel was obstructed at high relative humidity (RH), which enhanced SOA formation, but the opposite effect of RH was observed on SOA formation in the OH-oxidation channel. The different relative proportions of  $\text{O}_3$ - and OH-oxidation channels with different  $\text{NO}_2$  concentrations in the VOC photooxidation process might be a key factor underlying the inconsistencies in the effect of RH on SOA formation. Above all, the effect of  $\text{NO}_2$  on SOA formation with the change in other environmental factors is complex and should be clarified to further understand the mechanisms of SOA formation.

In the atmosphere of urban areas, aromatic compounds are the main VOCs, and they often dominate SOA formation (Zhang et al., 2019b; Gu et al., 2021). Therefore, in the present study, *o*-xylene was used to study the effect of  $\text{NO}_2$  on SOA formation under different RH conditions. This paper presents the evolution of SOA mass concentration, optical characteristics, and chemical components of SOA from *o*-xylene under controlled  $\text{NO}_2$  and RH conditions. Meanwhile, the relationships between the optical properties and chemical compositions of *o*-xylene derived SOA were also characterized.

## 2. Experimental methods

### 2.1. Chamber studies

All experiments were performed in a Teflon photooxidation chamber built in the East China Normal University. The description and characterization results of the chamber have been described in detail in previous studies (Liu et al., 2021b, 2021c), and only a brief introduction is provided here. A  $5 \text{ m}^3$  chamber made with the Teflon-FEP film (0.06 mm) was located in a temperature controlled stainless steel cube. UV-light lamps (TUV36W, Philips), which were used to drive OH radical formation through hydrogen peroxide ( $\text{H}_2\text{O}_2$ ) photolysis, surrounded the chamber. The light spectrum of UV-light had a peak intensity at 254 nm. Prior to each experiment, the chamber was initially evacuated and then filled with purified air for cleaning. The cycle of filling-purging was performed more than 3 times to reduce the concentrations of residual hydrocarbons,  $\text{O}_3$ , and  $\text{NO}_2$  to less than 1 ppbv, and the concentration of aerosols was less than  $10 \text{ m}^{-3}$ .

For the photooxidation experiment, the chamber was filled with zero air up to  $5 \text{ m}^3$ . Zero air was generated using a zero-air supply (111-D3N, Thermo Scientific™, USA), and it was humidified through introduction into a gas-washing bottle filled with high-purity water. The RH of zero air was controlled to create dry ( $\sim 20\%$  RH) or humid ( $\sim 70\%$  RH) conditions during different experiments. A known volume of liquid *o*-xylene (Sigma-Aldrich, analytically pure) was first injected into a Teflon FEP tube and then dispensed into the chamber with zero air. In the chamber, *o*-xylene concentrations were calculated were calculated from the injection volume of the liquids. Different volumes of  $\text{NO}_2$  (Air Liquid Shanghai, 510 ppm  $\text{NO}_2$  in  $\text{N}_2$ ) were introduced into the chamber directly from standard cylinders for the required concentrations. The initial conditions for different experiments are summarized in Table 1. The concentrations of *o*-xylene and  $\text{NO}_2$  in the experiments were higher than those in the real atmosphere to keep the particle production significant enough for the off-line collections and accurate measurements. The *o*-xylene concentrations remained stable in the different experimental conditions, the variations of *o*-xylene-derived SOA mass concentration and yield were only affected by the  $\text{NO}_2$  concentration in this study. The  $[\text{VOCs}]_0/[\text{NO}_2]_0$  ratios in this study covered the atmospheric

**Table 1**

Conditions for the *o*-xylene photooxidation experiment with different  $[\text{NO}_2]_0$  and RH.

Exp. No.	<i>o</i> -xylene	$\text{H}_2\text{O}_2$	$[\text{NO}_2]_0$	$[\text{VOCs}]_0/[\text{NO}_2]_0$	RH	SOA mass conc. $\mu\text{g m}^{-3}$	$\text{MAC}_\lambda = \frac{V_1}{V_a \cdot L \cdot M}$ $\text{m}^2 \text{g}^{-1}$
	$10^{13}$ molecule $\text{cm}^{-3}$		ppbv	ppbv/ppbv	%		
1	1.42	2.53	0	–	17	$54.2 \pm 1.8$	0.23
2			122	4.34	17	$74.3 \pm 1.3$	0.28
3			320	1.65	19	$96.3 \pm 2.1$	0.33
4			893	0.59	17	$127.2 \pm 0.6$	0.43
5		0	–	–	70	$177.7 \pm 2.4$	0.51
6		121	4.38	–	72	$169.5 \pm 1.4$	0.61
7		306	1.73	–	72	$153.8 \pm 1.2$	0.71
8		862	0.61	–	69	$146.5 \pm 1.5$	0.89

conditions from the clean to the polluted (Zou et al., 2015; Han et al., 2015), thus our experimental conditions are informative to aerosol formation in the real atmosphere. When all reactants were well mixed, the UV-light lamps were turned on and photooxidation was initiated.

## 2.2. Online instruments

Online instruments measured the  $\text{NO}_2$ ,  $\text{O}_3$ , and SOA concentrations continuously throughout the photooxidation process. The concentrations of  $\text{O}_3$  and  $\text{NO}_2$  were measured using an  $\text{O}_3$  analyzer (Model 49C, Thermo Electron Corporation, USA) and  $\text{NO}/\text{NO}_2/\text{NO}_2$  analyzer (Model 42C, Thermo Electron Corporation, USA), respectively. The formed SOA was measured with a scanning mobility particle sizer (SMPS), which was coupled with a differential mobility analyzer (DMA model 3081, TSI Inc., USA) for the measurement of particle size distributions and a condensation particle counter (CPC model 3776, TSI Inc., USA) for the measurement of concentrations. The sheath gas and sample gas flow rates were  $3 \text{ L min}^{-1}$  and  $0.3 \text{ L min}^{-1}$ , respectively, in the SMPS. The particle size scan ranged from 13.6 nm to 726.5 nm. A density of  $1.4 \text{ g cm}^{-3}$  for the aromatic hydrocarbon SOA was assumed when converting the volume concentration of SOA measured by the SMPS to mass concentration (Ng et al., 2007b; Liu et al., 2021b).

The chemical composition of SOA makes it very complicated to identify every photooxidation product. Hence, we used a high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS, Aerodyne Research Inc., USA) to detect the whole SOA chemical composition. It should be pointed out that only the particles with an aerodynamic equivalent diameter below  $1 \mu\text{m}$  could pass through the aerodynamic lens at the injection port and be detected by AMS. Because the W-mode of AMS produces high levels of noise while perform an analysis, only the V-mode of AMS was used in this study. The standard SQUIRREL 1.51H and PIKA 1.10H applications were retrieved from <http://cires1.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware> in Igor Pro (WaveMetrics, Inc., USA) for the HR-ToF-AMS data analysis. Notably, for comparability with previous studies, all bulk SOA elemental ratios (i. e. O/C, H/C, and N/C) and mass-to-carbon ratio (OM/OC) were calculated based on the method proposed by Aiken et al. (2008) and Canagaratna et al. (2015).

## 2.3. SOA collection

After 60 min of reaction, the mass concentration of SOA reached its maximum. The particles were collected on 47 mm

polytetrafluoroethylene (PTFE) filters ( $0.22 \mu\text{m}$  pore size) with a sampling volume of  $3 \text{ m}^3$ . PTFE filters were weighed before and after sampling to determine the amount of particles collected. The PTFE filter samples were extracted with 10 mL of methanol under ultrasonication for offline optical property characterization.

The light optical absorption of *o*-xylene derived SOA was analyzed with a UV-Vis spectrometer (P9, Shanghai Mapada, China). In order to eliminate the influence of the amount of SOA collected on the filter on absorbance, the mass absorption coefficient ( $\text{MAC}$ ,  $\text{m}^2 \text{g}^{-1}$ ) was used for comparison. The MAC value was calculated using Eq. (1) presented below:

$$\text{MAC}_\lambda = (A_\lambda - A_{700}) \cdot \frac{V_1}{V_a \cdot L \cdot M} \cdot \ln(10) \quad (1)$$

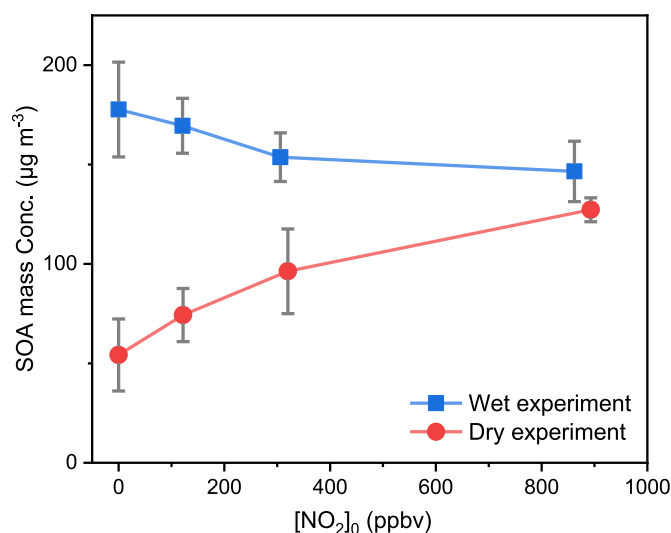
where,  $\text{MAC}_\lambda$  is the mass absorption coefficient of SOA at the  $\lambda$  wavelength;  $A_\lambda$  and  $A_{700}$  are the light absorption intensity at a specific wavelength and background value, respectively;  $V_1$  is the volume of methanol with dissolved aerosols;  $V_a$  is the volume of the sampled air;  $L$  (1 cm) is the optical path length; and  $M$  ( $\mu\text{g m}^{-3}$ ) is the mass concentration of methanol-soluble organic matter.

## 3. Results and discussion

### 3.1. SOA mass concentrations

The mass concentrations of *o*-xylene SOA formed with different initial  $\text{NO}_2$  concentrations and RH conditions are shown in Fig. 1. The particle wall loss rates were measured at the end of the chamber experiment after the UV-lamps were turned off, and the mass concentration was corrected with the same way of Jiang et al. (2020) and Pathak et al. (2007), which have been described in our previous studies (Liu et al., 2021b; Liu et al., 2022). It needs to be pointed out that the SOA formed in humid conditions was not dried before the SMPS measurement. In order to exclude the contributions of aerosol liquid water (ALW) in particles, a volume growth factor (VGF) of 1.3 was used for the calibration of SOA mass concentration. Here, VGF is defined as the ratio of the particle volume in a humid condition to that after drying (Engelhart et al., 2011). The SOA mass concentrations formed in humid conditions were calibrated based on the VGF in Fig. 1.

In dry conditions, the mass concentration of *o*-xylene derived SOA increased with the increasing initial  $\text{NO}_2$  concentration. The SOA mass concentration was  $54.2 \pm 1.8 \mu\text{g m}^{-3}$  in the absence of  $\text{NO}_2$ , and it increased to  $74.3 \pm 1.3 \mu\text{g m}^{-3}$ ,  $96.3 \pm 2.1 \mu\text{g m}^{-3}$ , and  $127.2 \pm 0.6 \mu\text{g m}^{-3}$



**Fig. 1.** Mass concentrations of SOA generated by photooxidation of *o*-xylene under different  $\text{NO}_2$  concentrations under dry and humid conditions.

$\text{m}^{-3}$  when the initial  $\text{NO}_2$  concentration was increased to 122 ppbv, 320 ppbv, and 893 ppbv, respectively. A nonlinear trend of an initial increase was followed by a decrease in SOA formation with increasing  $\text{NO}_2$  concentration has been widely observed in previous studies (Wildt et al., 2014; Zhao et al., 2018; Liu et al., 2021c). The reaction between NO and  $\text{RO}_2$ , which are formed from the OH-initiated oxidation, which produces higher volatility products, can suppress SOA formation in high  $\text{NO}_x$  conditions. However, in this study, the SOA formation was not inhibited even when the initial  $\text{NO}_x$  concentration was increased to the very high level of 893 ppbv. UV light with center wavelength of 254 nm was used to drive  $\text{H}_2\text{O}_2$  photolysis, while it cannot cause the photolysis of  $\text{NO}_2$  to NO as black light in the photooxidation process. In this study,  $\text{NO}_2$  was the main form of  $\text{NO}_x$ , and even in the experiment with high  $\text{NO}_x$  concentration, the reaction between  $\text{RO}_2$  and NO was not important and the suppression of SOA formation by  $\text{NO}_x$  was not observed. The reaction between OH and  $\text{NO}_2$  to form nitric acid is one of the most influential reactions in atmospheric chemistry (Möller et al., 2010). Liu et al. (2021c) reported that nitric acid can participate in the particle-phase, and observed that the concentration ratio of nitrate to SOA increased with increasing  $\text{NO}_x$  concentration through the  $\text{NO}_2 + \text{OH}$  reaction. More nitrate would therefore increase aerosol acidity. The semi-volatile organic compound products can readily partition into the particle phase through acid-catalyzed heterogeneous reactions via hydration, polymerization, and hemiacetal/acetate formation; thus, this results in the substantial promotion of SOA formation (Jang et al., 2002; Zhao et al., 2018). For different initial  $\text{NO}_2$  concentrations, the mass concentration ratio of nitrate ions to SOA ( $[\text{NO}_3^-]/[\text{Org}]$ ) was compared, as shown in Fig. 2. An increase in the ratio of  $[\text{NO}_3^-]/[\text{Org}]$ , which suggested a higher acidity of secondary aerosols, was observed with an increasing initial  $\text{NO}_2$  concentration. Hence, the enhanced acid-catalyzed reaction was responsible for the promotion of SOA formation with increased  $\text{NO}_2$  concentration.

Sarrafzadeh et al. (2016) suggested that an increase in SOA yield could arise from higher OH concentrations as they promote the formation of low volatility products in  $\beta$ -pinene photooxidation. SOA yield formed from artificial VOCs also appeared to have a strong correlation with OH concentrations, as shown in the studies by Healy et al. (2009) and Qi et al. (2020). A higher oxidation rate results in sufficiently high concentrations of photooxidation products to ensure that more photo-oxidation products are condensed in the particle-phase and promote SOA formation rather than remaining in the gas-phase, leading to further oxidation. In this study, the  $\text{H}_2\text{O}_2$  concentration remained constant in different experiments because the same volume of hydrogen peroxide solution was introduced into the chamber, and there was no significant difference in OH concentrations formed from  $\text{H}_2\text{O}_2$  photolysis. Nitrous acid (HONO) is another important OH precursor in the classical photooxidation process in the atmosphere. HONO could have been generated from the heterogeneous reaction of gas-phase  $\text{NO}_2$  on the chamber walls (Finlayson-Pitts et al., 2003; Ge et al., 2019). An

increase in the initial  $\text{NO}_2$  concentration would therefore be expected to result in higher OH concentration formed from the photolysis pathway of HONO, and promote SOA formation. Some previous studies have also pointed out that introducing additional HONO could greatly lead to producing higher SOA mass concentrations (Ng et al., 2007b; Healy et al., 2009). For the field observation and model simulation, SOA concentrations were also enhanced when OH radical-formation from the photolysis of HONO, which was generated by the heterogeneous reaction of  $\text{NO}_2$ , was considered (Liu et al., 2021a; Zhang et al., 2019a). The formation of HONO, which could then produce more OH through photolysis, was responsible for the enhancement of SOA mass concentration to a certain extent in the higher  $\text{NO}_2$  concentration conditions. However, Liu et al. (2021a) found that the ratio of HONO/ $\text{NO}_2$  was in the range of 4%–12% at the urban atmosphere, and the contribution of OH from the photolysis of HONO, which was formed through the heterogeneous reaction of  $\text{NO}_2$ , contributes to 3.4%–20% of the atmospheric OH. For this diversion ratio of HONO formation and photolysis, in this study, the heterogeneous conversion of  $\text{NO}_2$  would provide a maximum of 16 ppbv of OH, which was lower by more than an order of magnitude of the initial  $\text{H}_2\text{O}_2$  concentration. Considering the SOA mass concentration increased 1.3 times with the initial  $\text{NO}_2$  concentration increasing from 0 ppbv to about 900 ppbv, the formation of HONO from heterogeneous conversion of  $\text{NO}_2$  was not important for the enhancement of SOA formation when compared with the acid-catalyzed reaction in this study.

The *o*-xylene SOA mass concentrations formed in dry and humid conditions were also compared in this study. Fig. 1 shows that the mass concentrations of *o*-xylene derived SOA at humid conditions were much greater than those at dry conditions. The SOA mass concentrations in dry conditions ranged from  $54.2 \mu\text{g m}^{-3}$  to  $127.2 \mu\text{g m}^{-3}$ . But at high RH conditions, the *o*-xylene derived SOA mass concentration ranged from  $146.5 \mu\text{g m}^{-3}$  to  $177.7 \mu\text{g m}^{-3}$ . The maximum ratio of SOA mass concentration formed in humid conditions to that formed in dry conditions was 3.4, which occurred in the absence of  $\text{NO}_2$ . A positive effect of RH on SOA formation has also been obtained in other experiments (Kamens et al., 2011; Jia and Xu, 2018; Liu et al., 2019b). The ALW is a key component affecting SOA formation. The soluble oxidation products with high volatility, which are widely present in the gas-phase under dry conditions, were more concentrated on the surface of humid particulates, which increased the mass concentration of SOA (Liu et al., 2019b). Glyoxal and methylglyoxal are important oxidation products of *o*-xylene photooxidation in the atmosphere, but they did not make a significant contribution to SOA formation (only 1–3% for xylene derived SOA) (Nishino et al., 2010). Strong evidence has suggested that the gas phases of both glyoxal and methylglyoxal are uptaken by atmospheric particles, primarily by particle associated water (Volkamer et al., 2009). But Wang et al. (2016b) pointed out that dissolved glyoxal could not completely contribute to the increased mass concentration of SOA under high RH conditions based on Henry's Law constant (Wang et al., 2016b). Aqueous reactions could also have enhanced SOA formation. Dissolved glyoxal can equilibrate to geminal diol or hydrate forms, and rapidly dimerize or trimerize to form oligomers through acid catalyzed hemiacetal formation and aldol condensation (Loeffler et al., 2006; Corrigan et al., 2008; Jia and Xu, 2015; Peters et al., 2021). Hence, at high RH conditions, glyoxal and other soluble substances can produce additional organic compounds via aqueous photochemical reactions in ALW (White et al., 2014; Faust et al., 2017) which do not immediately transform back to the gas-phase, even when SOA is dried. De Haan et al. (2009) showed that more than 30% of the glyoxal mass was preserved in the particle phase when aerosol was dried from aqueous droplets. After the dissolved glyoxal or methylglyoxal reacts, more glyoxal or methylglyoxal was able to consistently participate in the particle phase. Even in the real atmosphere environment photochemical processing and aqueous-phase processing play an important role in SOA production in humid conditions. Based on the formation and evolution characteristics of SOA at urban area, Li et al. (2020) have pointed the enhanced aqueous-phase

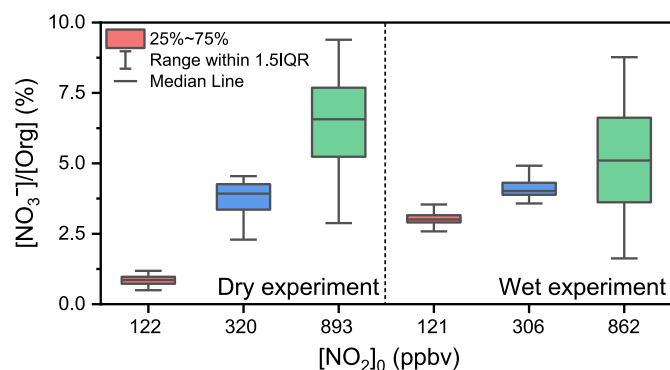


Fig. 2. Mass concentration ratio of nitrate ions to SOA ( $[\text{NO}_3^-]/[\text{Org}]$ ) for different experimental conditions.



oxidation and the partitioning process both contributed to the incorporation of oxygenated species into the particle phase in humid conditions and enhanced the SOA mass concentration, which is consistent with our result.

Although an increased RH can promote the formation of *o*-xylene derived SOA, this promoting effect gradually decreased with increasing NO<sub>2</sub> concentration. In contrast to the effect of NO<sub>2</sub> on SOA formation under dry conditions, the mass concentration of *o*-xylene SOA formed in humid conditions gradually decreased with increasing initial NO<sub>2</sub> concentration. As pointed out before, both liquid absorption and liquid phase heterogeneous reactions of glyoxal and methylglyoxal were responsible for increases in SOA mass concentration under high humidity conditions. Nishino et al. (2010) measured the formation yields of glyoxal and methylglyoxal from the gas-phase OH radical-initiated reactions with xylenes, and they found that the yields decreased with increased NO<sub>2</sub> concentration. Their experiment result showed that the glyoxal and methylglyoxal formation yields decreased from 14.3% to 11.5% and from 35.5% to 30.4%, respectively, with the initial NO<sub>2</sub> concentration increasing from 0 ppbv to 893 ppbv. Atkinson and Aschmann (1994) also investigated the effect of NO<sub>2</sub> concentration on the gas-phase products from aromatic hydrocarbons photooxidation, and found that both glyoxal and methylglyoxal decreased with increasing NO<sub>2</sub> concentration in xylene photooxidation process. The atmospheric model results also showed a decrease trend of glyoxal and methylglyoxal yield from the photooxidation of xylene with increasing NO<sub>2</sub> concentration (Bates et al., 2021). The decreased concentrations of glyoxal and methylglyoxal, which could be condensed on humid particles and caused an increase in the SOA mass concentration under high humidity conditions, was weakened with increasing NO<sub>2</sub> concentration. The above results indicated that the decreases in glyoxal and methylglyoxal formation directly caused the decrease in the *o*-xylene derived SOA mass concentration with increasing NO<sub>2</sub> concentrations in humid conditions.

The enhancement of the SOA aging processes at high RH conditions might be another reason that led to different trends in SOA formation with NO<sub>2</sub> under humid and dry conditions. Except for the gas phase reaction, OH radicals can react with particulate organic matter to form small-molecule volatile compounds, which can serve as a sink for SOA (Molina et al., 2004). The OH uptake coefficient is different between low and high RH conditions. In humid conditions, hydrogen bonds between LWC and OH radicals can enhance the attachment of OH radicals and moist SOA surfaces, which would ensure that more OH radicals could react with the condensed organic products (Chan et al., 2014). In addition to its effect on the reaction of OH radicals on the surfaces of SOA, the difference in viscosity and molecular diffusion capacity in the bulk particle phase under different RH conditions would affect the reactivity of OH towards the interiors of SOA. The influence of RH on particle phase viscosity plays an important role in the heterogeneous oxidation of organic aerosols. Moog et al. (1982) showed that the orientation stabilities of molecules in the particle phase were damaged, which could enhance the mobility and decrease the aerosol viscosity of SOA with increases in RH. Smith et al. (2021) also pointed out that the reactive uptake and diffusivity are kinetically inhibited in dry conditions for aromatic SOA because of the high viscosity, and the limitations in the mass transfer rate for uptake are gradually reduced with increasing RH. The heterogeneous reaction of SOA transitioned from diffusion-limited in dry conditions to saturated uptake in humid conditions. The viscosity of condensed organic species would relate to their diffusion coefficient ( $D_{org}$ ), and would impact the mixing timescale of SOA. The lower the viscosity, the shorter the mixing timescale. In general, in dry conditions,  $D_{org}$  is  $10^{-15}$  cm<sup>2</sup> s<sup>-1</sup> when the corresponding mixing timescale is about 40 min for a 100 nm particle. However,  $D_{org}$  increases to  $10^{-12}$  cm<sup>2</sup> s<sup>-1</sup> in humid conditions, and the mixing timescale for a 100 nm particle is decreased to about 2 s (Koop et al., 2011). An increase in RH could favor the diffusion of OH into the particles. Hence, the aging process of OH radicals may occur throughout the entire particle, not

only on the surface of the particle, which would cause an increase in the particle reactivity with increasing RH (Kuwata and Martin, 2012). More OH would react with the condensed organic species to produce small molecule products through fragmentation reactions. Heterogeneous OH oxidation of SOA under low- and high-RH conditions have been studied previously. The aerosol volume loss was more efficient in humid conditions compared to dry conditions (Romonosky et al., 2015; Liu et al., 2019a). The higher net fragmentation values in high-RH conditions were responsible for the loss of SOA through heterogeneous OH oxidation. Li et al. (2018) also pointed out that the aging process of SOA through OH oxidation was accelerated at high RH, and there was a substantial difference in the SOA chemical composition between high RH and low RH conditions. The above findings indicated that an increased mass loss rate of formed SOA occurred with increasing RH at constant OH concentrations. Meanwhile, increased OH concentrations might further increase the mass loss rate of formed SOA in high RH conditions. HONO formation through disproportionation of a heterogeneous reaction of gas-phase NO<sub>2</sub> was increased with increasing initial NO<sub>2</sub> concentration and RH. More OH formed through HONO photolysis, which further enhanced the SOA aging degree under high NO<sub>2</sub> concentrations in humid conditions (Forstner et al., 1997; Zhang et al., 2019b). Due to these reasons, a decreased SOA mass concentration was observed with an increase in NO<sub>2</sub> concentration in high RH conditions. It should be pointed out here that NO<sub>2</sub> even suppressed *o*-xylene derived SOA formation in humid conditions, the SOA mass concentration in humid conditions was still higher than that in dry conditions, even with high NO<sub>2</sub> levels.

### 3.2. SOA light absorption

Optical absorption is an important property for evaluating the influence of SOA on solar radiation and climate effects. The  $MAC_{\lambda=365\text{ nm}}$  of *o*-xylene derived SOA with different NO<sub>2</sub> concentrations and RH conditions was investigated in this section and is shown in Fig. 3.

As shown in Fig. 3, an increased  $MAC_{\lambda=365\text{ nm}}$  value increased with NO<sub>2</sub> concentration in both low and high RH conditions.  $MAC_{\lambda=365\text{ nm}}$  of the *o*-xylene derived SOA was  $0.23\text{ m}^2\text{ g}^{-1}$  in the absence of NO<sub>2</sub>, and it increased to  $0.43\text{ m}^2\text{ g}^{-1}$  when the initial concentration of NO<sub>2</sub> was increased to 893 ppbv in dry conditions. In humid conditions,  $MAC_{\lambda=365\text{ nm}}$  increased from  $0.51\text{ m}^2\text{ g}^{-1}$  to  $0.89\text{ m}^2\text{ g}^{-1}$  as the initial NO<sub>2</sub> concentration increased from 0 ppbv to 862 ppbv. The effect of increasing NO<sub>2</sub> on the SOA light-absorbing ability was in accordance with previous studies (Xie et al., 2017; Liu et al., 2022). However, the

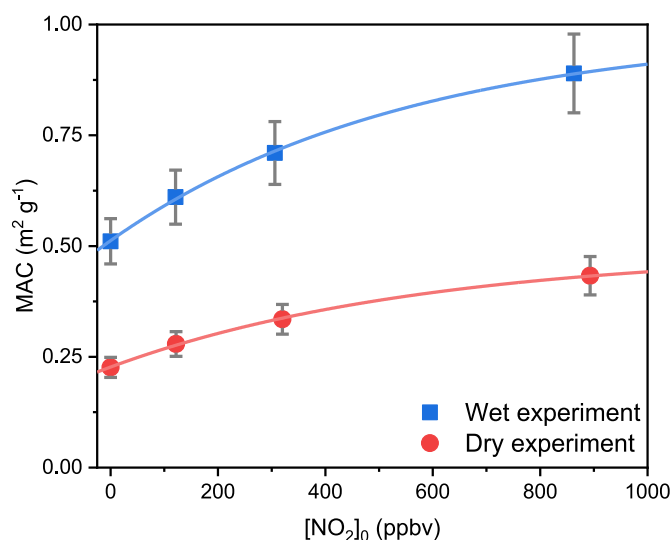


Fig. 3. Evolution of  $MAC_{\lambda=365\text{ nm}}$  of *o*-xylene SOA formed with different initial NO<sub>2</sub> concentrations at both humid and dry conditions.

$MAC_{\lambda=365\text{ nm}}$  of *o*-xylene derived SOA did not have a linear relationship with initial  $\text{NO}_2$  concentration. The increase in the ratio of  $MAC_{\lambda=365\text{ nm}}$  with  $\text{NO}_2$  gradually decreased with increasing  $\text{NO}_2$  concentration. In the absence of  $\text{NO}_2$ , the increase in the ratio of  $MAC_{\lambda=365\text{ nm}}$  with  $\text{NO}_2$  was  $4.4 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$  in dry conditions and  $9.0 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$  in humid conditions. In dry conditions, when the initial  $\text{NO}_2$  concentrations increased to 122 ppbv, 320 ppbv, and 893 ppbv, the increase in the ratio of  $MAC_{\lambda=365\text{ nm}}$  with  $\text{NO}_2$  was decreased to  $3.6 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$ ,  $2.7 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$ , and  $8.2 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$ , respectively. In humid conditions, when the initial  $\text{NO}_2$  concentration increased to 121 ppbv, 306 ppbv, and 862 ppbv, the increase in the ratio of  $MAC_{\lambda=365\text{ nm}}$  with  $\text{NO}_2$  was decreased to  $7.0 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$ ,  $4.8 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$ , and  $1.1 \times 10^{-4} \text{ m}^2 \text{ g}^{-1} \text{ ppbv}^{-1}$ , respectively.

In addition, for the experiment with similar initial  $\text{NO}_2$  concentrations, the  $MAC_{\lambda=365\text{ nm}}$  was higher for *o*-xylene SOA formed in humid conditions than that formed in dry conditions. The  $MAC_{\lambda=365\text{ nm}}$  of SOA increased 2.1–2.3 times from the low RH to high RH conditions. It is important to point out that, in dry conditions, even if the initial  $\text{NO}_2$  concentration was increased to 893 ppbv, which is almost impossibly high for real environments, the  $MAC_{\lambda=365\text{ nm}}$  of *o*-xylene derived SOA was  $0.43 \text{ m}^2 \text{ g}^{-1}$ , which was still lower than that formed in the absence of  $\text{NO}_2$  under a high RH condition ( $0.51 \text{ m}^2 \text{ g}^{-1}$ ). Clearly, RH is an important factor that strongly affects the absorbance of SOA.

### 3.3. SOA chemical composition

The chemical composition of SOA generated through OH oxidation of *o*-xylene with different RH and  $\text{NO}_2$  concentrations was investigated online with a HR-ToF-AMS. In previous photooxidation studies, most SOA formation mechanisms were obtained based on the chemical compositions, which were detected through offline mass spectrometry (Klodt et al., 2022; Shen et al., 2018; Romonosky et al., 2015). Relative changes in the SOA chemical composition occur through the photooxidation process, which is important in understanding of the mechanism of SOA formation, but is often neglected. The changes in mass spectra are considered to be a function of OH exposure. Here, the average carbon oxidation state ( $\text{OS}_C \approx 2 \text{ O/C-H/C}$ ) was used to describe the changes in the degree of oxidation of *o*-xylene derived SOA during the photooxidation process. The different  $\text{OS}_C$  values and change trends observed for *o*-xylene SOA formed in different  $\text{NO}_2$  concentrations in dry and humid conditions are shown in Fig. 4.

For photooxidation in dry condition, the  $\text{OS}_C$  of *o*-xylene derived SOA decreased with increasing initial  $\text{NO}_2$  concentration. This result suggested that some high volatility oxidation products with lower  $\text{OS}_C$  values, which were present in the gas phase in experiments without  $\text{NO}_2$ , was able to participate in the particle phase in the presence of  $\text{NO}_2$  through the acid-catalyzed reactions (Kroll et al., 2011). The lower  $\text{OS}_C$  value for photooxidation with a higher initial  $\text{NO}_2$  concentration also illustrated that the acid-catalyzed reaction was enhanced by increased

$\text{NO}_2$  concentrations, which proved that the promotion of SOA formation was attributed to the increase in acid-catalyzed reactions, as pointed out in Section 3.1.

When we compared the  $\text{OS}_C$  of SOA formed in dry and humid conditions in the absence of  $\text{NO}_2$ , the  $\text{OS}_C$  value was lower for SOA formed in humid conditions than in dry conditions. This result also suggested that some water-soluble compounds with lower  $\text{OS}_C$  and high volatility were present in the gas phase in dry conditions, which could increase the SOA mass concentration via uptake or aqueous reactions in aerosol water in high RH conditions (Ng et al., 2007b; Jia and Xu, 2018). In contrast to dry conditions, the  $\text{OS}_C$  value for *o*-xylene SOA did not change appreciably with different initial  $\text{NO}_2$  concentrations in humid conditions. However, the change rate of  $\text{OS}_C$  with time increased with increasing  $\text{NO}_2$  concentration in humid conditions. The  $\text{OS}_C$  change rates of SOA for experiments with initial  $\text{NO}_2$  concentrations of 121 ppbv, 306 ppbv, and 862 ppbv were increased by 12.8%, 20.6%, and 59.1%, respectively, relative to when  $\text{NO}_2$  was absent in humid conditions. The rate of change of  $\text{OS}_C$  reflects the aging rate of SOA (Han et al., 2016). Thus, the aging process of *o*-xylene derived SOA was promoted by increases in the  $\text{NO}_2$  concentration in humid conditions, and the enhanced aging process was mainly responsible for the suppression of SOA formation by  $\text{NO}_2$  in humid conditions. Although an increased  $\text{NO}_2$  concentration can promote OH concentration by increasing the formation of HONO, the  $\text{OS}_C$  change rate of SOA in experiments in dry condition was not affected by increases in the  $\text{NO}_2$  concentration. Both OH concentration and reactive activation of OH with SOA were key factors affecting SOA aging. The different  $\text{OS}_C$  change rates of SOA with different initial  $\text{NO}_2$  concentrations in dry or humid conditions also illustrated that the aging process of SOA was restricted in dry conditions because of the low mobility and high viscosity of SOA, which inhibited the diffusion of OH into SOA and the aging reaction of OH with the internal organic components of SOA. In humid conditions, the mobility of SOA was not the limiting factor for OH reactions with organic components of SOA, as pointed out in Section 3.2, and the aging process of *o*-xylene derived SOA was directly affected by the OH concentration. The above results indicated that the aging process of SOA with increases in the OH concentration was not consistent between dry and humid conditions, and the aging process was faster in high RH conditions.

It has been pointed out that NOCs can be formed through photooxidation in the presence of  $\text{NO}_2$ , and NOCs are important light-absorbing components in SOA (Lee et al., 2014; Laskin et al., 2015; Li et al., 2021). The relationship between  $MAC_{\lambda=365\text{ nm}}$  and the content of NOC fragments in the total SOA mass under different  $\text{NO}_2$  concentrations in both dry and humid conditions is shown in Fig. 5. At similar  $\text{NO}_2$  concentrations, the NOC content of *o*-xylene SOA formed in humid conditions was higher than that formed in dry conditions. This shows that an increased RH is beneficial to the formation of nitrogenous organic matter in SOA. Li et al. (2021) analyzed light absorption and chromophore properties of SOA in the winter of Beijing, and they indicated the aqueous-phase pathways were more important for elevated formation of

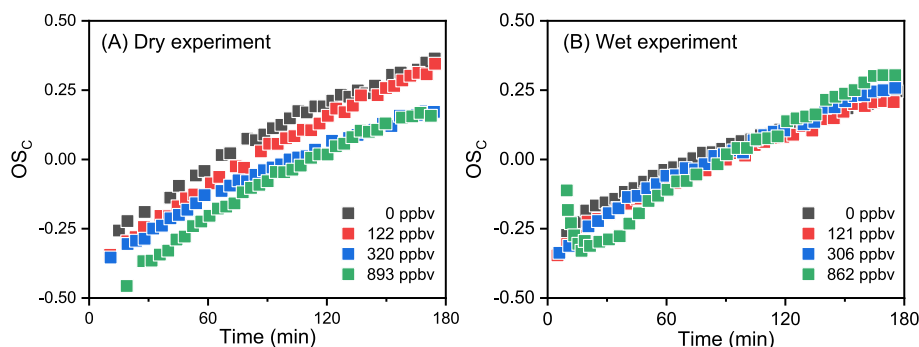


Fig. 4. Evolution of  $\text{OS}_C$  values for *o*-xylene SOA formed at different initial  $\text{NO}_2$  concentrations in dry (A) and humid (B) conditions.

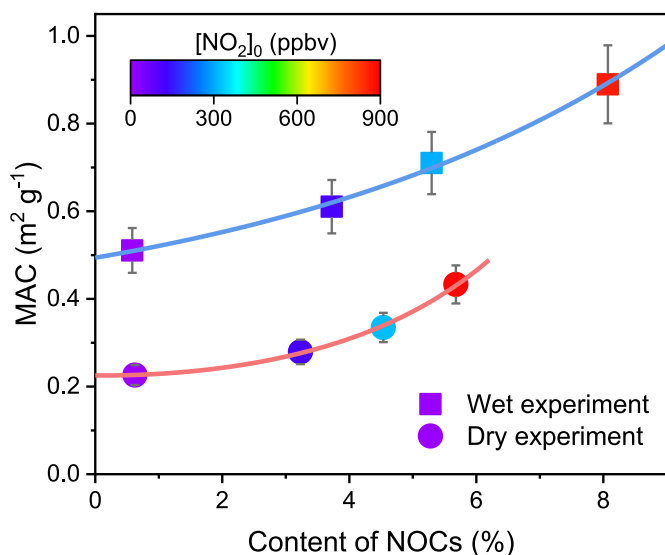


Fig. 5. Relationship between  $MAC_{\lambda = 365 \text{ nm}}$  and NOCs of *o*-xylene SOA under different  $[NO_2]_0$  conditions in both dry and humid conditions.

NOCs under a higher RH condition. Previous chamber studies also pointed out that more NOCs were formed through atmospheric aqueous phases in the presence of  $NO_2$  (Smith et al., 2016; He et al., 2019). The increased NOC content with increased RH observed here was in agreement with those field and chamber studies, which indicated that the atmospheric aqueous phase reactions are more important than gas phase reactions for NOC formation and increase the content of NOCs in SOA. Meanwhile, considering that the SOA mass concentration was higher in humid conditions than in dry conditions, the increase in the amount of NOCs was even more significant in high RH conditions. When compared with dry conditions, the NOC content of *o*-xylene SOA formed in humid conditions was higher by 15.3%, 16.8%, and 42.1% with initial  $NO_2$  concentrations of about 120 ppbv, 320 ppbv, and 900 ppbv, respectively. However, the amount of NOC fragments increased by 45.2%, 34.8%, and 63.0% in humid conditions, more than in dry conditions, in the experiments with initial  $NO_2$  concentrations of about 120 ppbv, 320 ppbv, and 900 ppbv, respectively.

Fig. 5 shows that the  $MAC_{\lambda = 365 \text{ nm}}$  of *o*-xylene derived SOA increased along with the content of NOCs in both dry and humid conditions, but  $MAC_{\lambda = 365 \text{ nm}}$  was not linearly related with the NOC content of *o*-xylene SOA. The ratio of  $MAC_{\lambda = 365 \text{ nm}}$  to NOC content increased with increasing initial  $NO_2$  concentration, and the enhancement of this ratio was more obvious in dry conditions than humid conditions. When comparing between dry and humid conditions, the  $MAC_{\lambda = 365 \text{ nm}}$  of SOA formed in humid conditions was much higher than that in dry conditions. Based on the above findings, it appeared that the content of NOCs is not the only factor that affects the light absorption of SOA. Other types of oxidation products could also increase the absorption of SOA in humid and high  $NO_2$  conditions.

Humic-like substances (HULIS) are important light-absorbing substances in SOA (Hoffer et al., 2006; Laskin et al., 2015; Wu et al., 2022). Hoffer et al. (2006) also pointed out that HULIS could partially account for the unexplained fraction of light absorption observed in the troposphere. HULIS can be produced through aqueous processing, oligomerization of water-soluble organics, and multiphase chemistry reactions of organic products derived from photooxidation of anthropogenic VOCs (Herckes et al., 2013; Paglione et al., 2014). Zhang et al. (2020) also found that an increased RH could enhance the optical properties of SOA formed from OH oxidation of aromatic VOCs, and the formation of oligomers through liquid phase reactions might be responsible for the enhancement of light absorption in humid conditions. Based on the above findings, we suspected that the increased ratio of  $MAC_{\lambda = 365 \text{ nm}}$  to

NOC content at a high  $NO_2$  concentration was caused by HULIS formation through a heterogeneous acid-catalyzed reaction, and enhanced HULIS formation through a liquid reaction might be responsible for the higher  $MAC_{\lambda = 365 \text{ nm}}$  of *o*-xylene SOA formed in humid conditions with the same NOC content. However, the reaction mechanism and formation yield of HULIS from aromatic hydrocarbon was not considered in this study and should be explored in the future.

#### 4. Conclusion

The mass concentration and light absorption of *o*-xylene SOA formed from OH photooxidation with different initial  $NO_2$  concentrations were studied in both humid and dry conditions using a chamber simulation technique. When compared with the dry condition, the mass concentration of *o*-xylene derived SOA was enhanced in humid conditions. In the absence of  $NO_2$ , *o*-xylene SOA was higher by 3.4 times in the SOA formed in humid conditions compared to that formed in dry conditions. The increased SOA mass concentration under humid conditions was mainly due to dissolution of water-soluble products (e.g. glyoxal and methylglyoxal) and aqueous reactions. This study observed that  $NO_2$  had two opposite effects on SOA formation depending on whether conditions were dry or humid. When the initial  $NO_2$  concentration was increased from 0 ppbv to about 900 ppbv, *o*-xylene SOA mass concentration increased from  $54.2 \mu\text{g m}^{-3}$  to  $127.2 \mu\text{g m}^{-3}$  in dry conditions, but decreased from  $230.7 \mu\text{g m}^{-3}$  to  $190.3 \mu\text{g m}^{-3}$  in humid conditions. The AMS results showed that lower oxidation products participated in the particle phase through acid-catalyzed heterogeneous reactions resulting in the promotion of SOA formation at a high  $NO_2$  concentration in dry conditions, however, an increase in the aging process was responsible for the decreased SOA mass concentration with elevated  $NO_2$  in humid conditions.

Light absorption of *o*-xylene SOA was enhanced with increased  $NO_2$  concentration and RH conditions. The highest  $MAC_{\lambda = 365 \text{ nm}}$  of  $0.89 \text{ m}^2 \text{ g}^{-1}$  was obtained in humid conditions with an initial  $NO_2$  concentration of 862 ppbv, and was about 4 times higher than that obtained in dry conditions without  $NO_2$ . The content of NOC fragments also increased with increases in  $NO_2$  concentrations and RH, but it did not show a robust linear correlation with  $MAC_{\lambda = 365 \text{ nm}}$ . This result suggested that NOCs make an important contribution to light absorption of *o*-xylene SOA. But beyond that, we suspected that HULIS formed through heterogeneous acid-catalyzed reactions at a high  $NO_2$  concentration and liquid reactions contributed to the extra light absorption of *o*-xylene SOA formed at high  $NO_2$  and RH conditions.

#### Author contributions

SL and GW designed the experiment. SL, YW and XX conducted the experiments. SL, YW, and GW performed the data interpretation. SL and GW wrote the paper.

#### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

#### Data availability

No data was used for the research described in the article.

#### Acknowledgments

This work was financially supported by the National Natural Science Foundation of China (No. 42130704, 42005088), the China Postdoctoral Science Foundation (No. 2022T150215), Shanghai Science and Technology Innovation Action Plan (20dz1204000) and ECNU Happiness



Flower program.

## References

- Aiken, A.C., Decarlo, P.F., Kroll, J.H., Worsnop, D.R., Huffman, J.A., Docherty, K.S., Ulbrich, I.M., Mohr, C., Kimmel, J.R., Sueper, D., Sun, Y., Zhang, Q., Trimborn, A., Northway, M., Ziemann, P.J., Canagaratna, M.R., Onasch, T.B., Alfarra, M.R., Prevot, A.S., Dommen, J., Duplissy, J., Metzger, A., Baltensperger, U., Jimenez, J.L., 2008. O/C and OM/OC ratios of primary, secondary, and ambient organic aerosols with high-resolution time-of-flight aerosol mass spectrometry. *Environ. Sci. Technol.* 42, 4478–4485. <https://doi.org/10.1021/es703009q>.
- Aruffo, E., Wang, J., Ye, J., Ohno, P., Qin, Y., Stewart, M., McKinney, K., Di Carlo, P., Martin, S.T., 2022. Partitioning of organonitrates in the production of secondary organic aerosols from  $\alpha$ -pinene photo-oxidation. *Environ. Sci. Technol.* <https://doi.org/10.1021/acs.est.1c08380>.
- Atkinson, R., Aschmann, S.M., 1994. Products of the gas-phase reactions of aromatic hydrocarbons: effect of  $\text{NO}_2$  concentration. *Int. J. Chem. Kinet.* 26, 929–944. <https://doi.org/10.1002/kin.550260907>.
- Bates, K.H., Jacob, D.J., Li, K., Ivatt, P.D., Evans, M.J., Yan, Y., Lin, J., 2021. Development and evaluation of a new compact mechanism for aromatic oxidation in atmospheric models. *Atmos. Chem. Phys.* 21, 18351–18374. <https://doi.org/10.5194/acp-21-18351-2021>.
- Bianchi, F., Kurten, T., Riva, M., Mohr, C., Rissanen, M.P., Roldin, P., Berndt, T., Crounse, J.D., Wennberg, P.O., Mentel, T.F., Wildt, J., Junninen, H., Jokinen, T., Kulmala, M., Worsnop, D.R., Thornton, J.A., Donahue, N., Kjaergaard, H.G., Ehn, M., 2019. Highly oxygenated organic molecules (HOM) from gas-phase autoxidation involving peroxy radicals: a key contributor to atmospheric aerosol. *Chem. Rev.* 119, 3472–3509. <https://doi.org/10.1021/acs.chemrev.8b00395>.
- Canagaratna, M.R., Jimenez, J.L., Kroll, J.H., Chen, Q., Kessler, S.H., Massoli, P., Hildebrandt Ruiz, L., Fortner, E., Williams, L.R., Wilson, K.R., Surratt, J.D., Donahue, N.M., Jayne, J.T., Worsnop, D.R., 2015. Elemental ratio measurements of organic compounds using aerosol mass spectrometry: characterization, improved calibration, and implications. *Atmos. Chem. Phys.* 15, 253–272. <https://doi.org/10.5194/acp-15-253-2015>.
- Chan, M.N., Zhang, H.F., Goldstein, A.H., Wilson, K.R., 2014. Role of water and phase in the heterogeneous oxidation of solid and aqueous succinic acid aerosol by hydroxyl radicals. *J. Phys. Chem. C* 118, 28978–28992. <https://doi.org/10.1021/jp5012022>.
- Charan, S.M., Huang, Y., Seinfeld, J.H., 2019. Computational simulation of secondary organic aerosol formation in laboratory chambers. *Chem. Rev.* 119, 11912–11944. <https://doi.org/10.1021/acs.chemrev.9b00358>.
- Chu, B.W., Zhang, X., Liu, Y.C., He, H., Sun, Y., Jiang, J.K., Li, J.H., Hao, J.M., 2016. Synergetic formation of secondary inorganic and organic aerosol: effect of  $\text{SO}_2$  and  $\text{NH}_3$  on particle formation and growth. *Atmos. Chem. Phys.* 16, 14219–14230. <https://doi.org/10.5194/acp-16-14219-2016>.
- Corrigan, A.L., Hanley, S.W., De Haan, D.O., 2008. Uptake of glyoxal by organic and inorganic aerosol. *Environ. Sci. Technol.* 42, 4428–4433. <https://doi.org/10.1021/es7032394>.
- De Haan, D.O., Corrigan, A.L., Tolbert, M.A., Jimenez, J.L., Wood, S.E., Turley, J.J., 2009. Secondary organic aerosol formation by self-reactions of methylglyoxal and glyoxal in evaporating droplets. *Environ. Sci. Technol.* 43, 8184–8190. <https://doi.org/10.1021/es902152t>.
- Eddingsaas, N.C., Loza, C.L., Yee, L.D., Chan, M., Schilling, K.A., Chhabra, P.S., Seinfeld, J.H., Wennberg, P.O., 2012.  $\alpha$ -pinene photooxidation under controlled chemical conditions - Part 2: SOA yield and composition in low- and high-NOx environments. *Atmos. Chem. Phys.* 12, 7413–7427. <https://doi.org/10.5194/acp-12-7413-2012>.
- Engelhart, G.J., Hildebrandt, L., Kostenidou, E., Mihalopoulos, N., Donahue, N.M., Pandis, S.N., 2011. Water content of aged aerosol. *Atmos. Chem. Phys.* 11, 911–920. <https://doi.org/10.5194/acp-11-911-2011>.
- Faust, A., J., Wong P., J., Lee K., A., Abbatt P., J., 2017. Role of aerosol liquid water in secondary organic aerosol formation from volatile organic compounds. *Environ. Sci. Technol.* 51, 1405–1413. <https://doi.org/10.1021/acs.est.6b04700>.
- Finlayson-Pitts, B.J., Wingen, L.M., Sumner, A.L., Syomin, D., Ramazan, K.A., 2003. The heterogeneous hydrolysis of  $\text{NO}_2$  in laboratory systems and in outdoor and indoor atmospheres: an integrated mechanism. *Phys. Chem. Chem. Phys.* 5, 223–242. <https://doi.org/10.1039/b208564j>.
- Forstner, H.J.L., Flagan, R.C., Seinfeld, J.H., 1997. Secondary organic aerosol from the photooxidation of aromatic hydrocarbons: molecular composition. *Environ. Sci. Technol.* 31, 1345–1358. <https://doi.org/10.1021/es9605376>.
- Gaubert, B., Bouarar, I., Doumbia, T., Liu, Y., Stavrakou, T., Deroubaix, A., Darras, S., Elguindi, N., Granier, C., Lacey, F., Muller, J.F., Shi, X., Tilmes, S., Wang, T., Brasseur, G.P., 2021. Global changes in secondary atmospheric pollutants during the 2020 COVID-19 pandemic. *J. Geophys. Res. Atmos.* 126, e2020JD034213 <https://doi.org/10.1029/2020JD034213>.
- Ge, S.S., Wang, G.H., Zhang, S., Li, D.P., Xie, Y.N., Wu, C., Yuan, Q., Chen, J., Zhang, H., 2019. Abundant  $\text{NH}_3$  in China enhances atmospheric HONO production by promoting the heterogeneous reaction of  $\text{SO}_2$  with  $\text{NO}_2$ . *Environ. Sci. Technol.* 53, 14339–14347. <https://doi.org/10.1021/acs.est.9b04196>.
- George, C., Ammann, M., D'Anna, B., Donaldson, D.J., Nizkorodov, S.A., 2015. Heterogeneous photochemistry in the atmosphere. *Chem. Rev.* 115, 4218–4258. <https://doi.org/10.1021/cr500648z>.
- Gu, S., Guenther, A., Faiola, C., 2021. Effects of anthropogenic and biogenic volatile organic compounds on Los Angeles air quality. *Environ. Sci. Technol.* 55, 12191–12201. <https://doi.org/10.1021/acs.est.1c01481>.
- Han, T., Qiao, L., Zhou, M., Qu, Y., Du, J., Liu, X., Lou, S., Chen, C., Wang, H., Zhang, F., Yu, Q., Wu, Q., 2015. Chemical and optical properties of aerosols and their interrelationship in winter in the megacity Shanghai of China. *J. Environ. Sci.* 27, 59–69. <https://doi.org/10.1016/j.jes.2014.04.018>.
- Han, Z.W., Xie, Z.X., Wang, G.H., Zhang, R.J., Tao, J., 2016. Modeling organic aerosols over east China using a volatility basis-set approach with aging mechanism in a regional air quality model. *Atmos. Environ.* 124, 186–198. <https://doi.org/10.1016/j.atmosenv.2015.05.045>.
- He, L., Schaefer, T., Otto, T., Kroflic, A., Herrmann, H., 2019. Kinetic and theoretical study of the atmospheric aqueous-phase reactions of OH radicals with methoxyphenolic compounds. *J. Phys. Chem. A* 123, 7828–7838. <https://doi.org/10.1021/acs.jpca.9b05696>.
- Healy, R.M., Temime, B., Kuprovskiyte, K., Wenger, J.C., 2009. Effect of relative humidity on gas/particle partitioning and aerosol mass yield in the photooxidation of *p*-xylene. *Environ. Sci. Technol.* 43, 1884–1889. <https://doi.org/10.1021/es802404z>.
- Herckes, P., Valsaraj, K.T., Collett, J.L., 2013. A review of observations of organic matter in fogs and clouds: origin, processing and fate. *Atmos. Res.* 132–133, 434–449. <https://doi.org/10.1016/j.atmosres.2013.06.005>.
- Hoffer, A., Gelencsér, A., Guyon, P., Kiss, G., Schmid, O., Frank, G.P., Artaxo, P., Andreae, M.O., 2006. Optical properties of humic-like substances (HULIS) in biomass-burning aerosols. *Atmos. Chem. Phys.* 6, 3563–3570. <https://doi.org/10.5194/acp-6-3563-2006>.
- Huang, X.F., Dai, J., Zhu, Q., Yu, K., Du, K., 2020. Abundant biogenic oxygenated organic aerosol in atmospheric coarse particles: plausible sources and atmospheric implications. *Environ. Sci. Technol.* 54, 1425–1430. <https://doi.org/10.1021/acs.est.9b06311>.
- Jang, M., Czoschke, N.M., Lee, S., Kamens, R.M., 2002. Heterogeneous atmospheric aerosol production by acid-catalyzed particle-phase reactions. *Science* 298, 814–817. <https://doi.org/10.1126/science.1075798>.
- Jia, L., Xu, Y.F., 2015. Ozone and secondary organic aerosol formation from Ethylene- $\text{NO}_x$ -NaCl irradiations under different relative humidity conditions. *J. Environ. Sci.* 73, 81–100. <https://doi.org/10.1007/s10874-015-9317-1>.
- Jia, L., Xu, Y.F., 2018. Different roles of water in secondary organic aerosol formation from toluene and isoprene. *Atmos. Chem. Phys.* 18, 8137–8154. <https://doi.org/10.5194/acp-18-8137-2018>.
- Jiang, X.T., Lv, C., You, B., Liu, Z.Y., Wang, X.F., Du, L., 2020. Joint impact of atmospheric  $\text{SO}_2$  and  $\text{NH}_3$  on the formation of nanoparticles from photo-oxidation of a typical biomass burning compound. *Environ. Sci. Nano* 7, 2532–2545. <https://doi.org/10.1039/d0en00520g>.
- Kamens, R.M., Zhang, H., Chen, E.H., Zhou, Y., Parikh, H.M., Wilson, R.L., Galloway, K. E., Rosen, E.P., 2011. Secondary organic aerosol formation from toluene in an atmospheric hydrocarbon mixture: water and particle seed effects. *Atmos. Environ.* 45, 2324–2334. <https://doi.org/10.1016/j.atmosenv.2010.11.007>.
- Klodt, A.L., Adamek, M., Dibley, M., Nizkorodov, S.A., O'Brien, R.E., 2022. Effects of the sample matrix on the photobleaching and photodegradation of toluene-derived secondary organic aerosol compounds. *Atmos. Chem. Phys.* 22, 10155–10171. <https://doi.org/10.5194/acp-22-10155-2022>.
- Koop, T., Bookhold, J., Shiraiwa, M., Poschl, U., 2011. Glass transition and phase state of organic compounds: dependency on molecular properties and implications for secondary organic aerosols in the atmosphere. *Phys. Chem. Chem. Phys.* 13, 19238–19255. <https://doi.org/10.1039/c1cp22617g>.
- Kroll, J.H., Donahue, N.M., Jimenez, J.L., Kessler, S.H., Canagaratna, M.R., Wilson, K.R., Altieri, K.E., Mazzoleni, L.R., Wozniak, A.S., Bluhm, H., Mysak, E.R., Smith, J.D., Kolb, C.E., Worsnop, D.R., 2011. Carbon oxidation state as a metric for describing the chemistry of atmospheric organic aerosol. *Nat. Chem.* 3, 133–139. <https://doi.org/10.1038/nchem.948>.
- Kuwata, M., Martin, S.T., 2012. Phase of atmospheric secondary organic material affects its reactivity. *Proc. Natl. Acad. Sci. U.S.A.* 109, 17354–17359. <https://doi.org/10.1073/pnas.1209071109>.
- Laskin, A., Laskin, J., Nizkorodov, S.A., 2015. Chemistry of atmospheric brown carbon. *Chem. Rev.* 115, 4335–4382. <https://doi.org/10.1021/cr5006167>.
- Lee, H.J., Aiona, P.K., Laskin, A., Laskin, J., Nizkorodov, S.A., 2014. Effect of solar radiation on the optical properties and molecular composition of laboratory proxies of atmospheric brown carbon. *Environ. Sci. Technol.* 48, 10217–10226. <https://doi.org/10.1021/es502515r>.
- Li, J.Y., Liu, Z.R., Gao, W.K., Tang, G.Q., Hu, B., Ma, Z.Q., Wang, Y.S., 2020. Insight into the formation and evolution of secondary organic aerosol in the megacity of Beijing, China. *Atmos. Environ.* 220, 117070 <https://doi.org/10.1016/j.atmosenv.2019.117070>. ARTN 117070.
- Li, Q., Jiang, J., Afreh, I.K., Barsanti, K.C., Cocker Iii, D.R., 2022. Secondary organic aerosol formation from camphene oxidation: measurements and modeling. *Atmos. Chem. Phys.* 22, 3131–3147. <https://doi.org/10.5194/acp-22-3131-2022>.
- Li, X.R., Zhao, Q., Yang, Y., Zhao, Z.Y., Liu, Z.R., Wen, T.X., Hu, B., Wang, Y.S., Wang, L. L., Wang, G.H., 2021. Composition and sources of brown carbon aerosols in megacity Beijing during the winter of 2016. *Atmos. Res.* 262 <https://doi.org/10.1016/j.atmosres.2021.105773>. ARTN 105773.
- Li, Z., Smith, K.A., Cappa, C.D., 2018. Influence of relative humidity on the heterogeneous oxidation of secondary organic aerosol. *Atmos. Chem. Phys.* 18, 14585–14608. <https://doi.org/10.5194/acp-18-14585-2018>.
- Liu, J.Y., Liu, Z.R., Ma, Z.Q., Yang, S., Yao, D., Zhao, S., Hu, B., Tang, G., Sun, J., Cheng, M.T., Xu, Z.J., Wang, Y.S., 2021a. Detailed budget analysis of HONO in Beijing, China: implication on atmosphere oxidation capacity in polluted megacity. *Atmos. Environ.* 244, 117957 <https://doi.org/10.1016/j.atmosenv.2020.117957>.
- Liu, S.J., Jiang, X.T., Tsou, N.T., Lv, C., Du, L., 2019a. Effects of  $\text{NO}_x$ ,  $\text{SO}_2$  and RH on the SOA formation from cyclohexene photooxidation. *Chemosphere* 216, 794–804. <https://doi.org/10.1016/j.chemosphere.2018.10.180>.



- Liu, S.J., Tsona, N.T., Zhang, Q., Jia, L., Xu, Y.F., Du, L., 2019b. Influence of relative humidity on cyclohexene SOA formation from OH photooxidation. *Chemosphere* 231, 478–486. <https://doi.org/10.1016/j.chemosphere.2019.05.131>.
- Liu, S.J., Huang, D.D., Wang, Y.Q., Zhang, S., Liu, X.D., Wu, C., Du, W., Wang, G.H., 2021b. Synergetic effects of  $\text{NH}_3$  and  $\text{NO}_x$  on the production and optical absorption of secondary organic aerosol formation from toluene photooxidation. *Atmos. Chem. Phys.* 21, 17759–17773. <https://doi.org/10.5194/acp-21-17759-2021>.
- Liu, S.J., Wang, Y.Q., Wang, G.H., Zhang, S., Li, D.P., Du, L., Wu, C., Du, W., Ge, S.S., 2021c. Enhancing effect of  $\text{NO}_2$  on the formation of light-absorbing secondary organic aerosols from toluene photooxidation. *Sci. Total Environ.* 794, 148714 <https://doi.org/10.1016/j.scitotenv.2021.148714>.
- Liu, S.J., Liu, X., Wang, Y., Zhang, S., Wu, C., Du, W., Wang, G., 2022. Effect of  $\text{NO}_x$  and RH on the secondary organic aerosol formation from toluene photooxidation. *J. Environ. Sci.* 114, 1–9. <https://doi.org/10.1016/j.jes.2021.06.017>.
- Loeffler, K.W., Koehler, C.A., Paul, N.M., De Haan, D.O., 2006. Oligomer formation in evaporating aqueous glyoxal and methyl glyoxal solutions. *Environ. Sci. Technol.* 40, 6318. <https://doi.org/10.1021/es060810w>, 6123.
- Molina, M.J., Ivanov, A.V., Trakhtenberg, S., Molina, L.T., 2004. Atmospheric evolution of organic aerosol. *Geophys. Res. Lett.* 31 <https://doi.org/10.1029/2004gl020910>. Artn L22104.
- Mollner, A.K., Valluvadasan, S., Feng, L., Sprague, M.K., Okumura, M., Milligan, D.B., Bloss, W.J., Sander, S.P., Martien, P.T., Harley, R.A., McCoy, A.B., Carter, W.P., 2010. Rate of gas phase association of hydroxyl radical and nitrogen dioxide. *Science* 330, 646–649. <https://doi.org/10.1126/science.1193030>.
- Moog, R.S., Ediger, M.D., Boxer, S.G., Fayer, M.D., 1982. Viscosity dependence of the rotational reorientation of rhodamine B in mono- and polyalcohols. Picosecond transient grating experiments. *J. Phys. Chem.* 86, 4694–4700. <https://doi.org/10.1021/j100221a011>.
- Ng, N.L., Chhabra, P.S., Chan, A.W.H., Surratt, J.D., Kroll, J.H., Kwan, A.J., McCabe, D. C., Wennberg, P.O., Sorooshian, A., Murphy, S.M., Dalleska, N.F., Flagan, R.C., Seinfeld, J.H., 2007a. Effect of  $\text{NO}_x$  level on secondary organic aerosol (SOA) formation from the photooxidation of terpenes. *Atmos. Chem. Phys.* 7, 5159–5174. <https://doi.org/10.5194/acp-7-5159-2007>.
- Ng, N.L., Kroll, J.H., Chan, A.W.H., Chhabra, P.S., Flagan, R.C., Seinfeld, J.H., 2007b. Secondary organic aerosol formation from *m*-xylene, toluene, and benzene. *Atmos. Chem. Phys.* 7, 3909–3922. <https://doi.org/10.5194/acp-7-3909-2007>.
- Nishino, N., Arey, J., Atkinson, R., 2010. Formation yields of glyoxal and methylglyoxal from the gas-phase OH radical-initiated reactions of toluene, xylenes, and trimethylbenzenes as a function of  $\text{NO}_2$  concentration. *J. Phys. Chem. A* 114, 10140–10147. <https://doi.org/10.1021/jp105112h>.
- Noziere, B., Kalberer, M., Claeys, M., Allan, J., D'Anna, B., Decesari, S., Finessi, E., Glasius, M., Grgic, I., Hamilton, J.F., Hoffmann, T., Iinuma, Y., Jaoui, M., Kahnt, A., Kampf, C.J., Kourchev, I., Maenhaut, W., Marsden, N., Saarikoski, S., Schnelle-Kreis, J., Surratt, J.D., Szidat, S., Szmigielski, R., Wisthaler, A., 2015. The molecular identification of organic compounds in the atmosphere: state of the art and challenges. *Chem. Rev.* 115, 3919–3983. <https://doi.org/10.1021/cr5003485>.
- Offenberg, J.H., Lewandowski, M., Edney, E.O., Kleindienst, T.E., Jaoui, M., 2009. Influence of aerosol acidity on the formation of secondary organic aerosol from biogenic precursor hydrocarbons. *Environ. Sci. Technol.* 43, 7742–7747. <https://doi.org/10.1021/es901538e>.
- Paglion, M., Kiendler-Scharr, A., Mensah, A.A., Finessi, E., Giulianelli, L., Sandrini, S., Facchini, M.C., Fuzzi, S., Schlag, P., Piazzalunga, A., Tagliavini, E., Henzing, J.S., Decesari, S., 2014. Identification of humic-like substances (HULIS) in oxygenated organic aerosols using NMR and AMS factor analyses and liquid chromatographic techniques. *Atmos. Chem. Phys.* 14, 25–45. <https://doi.org/10.5194/acp-14-25-2014>.
- Pathak, R.K., Stanier, C.O., Donahue, N.M., Pandis, S.N., 2007. Ozonolysis of alpha-pinene at atmospherically relevant concentrations: temperature dependence of aerosol mass fractions (yields). *J. Geophys. Res. Atmos.* 112 <https://doi.org/10.1029/2006jd007436>. Artn D03201.
- Peters, J.H., Dette, H.P., Koop, T., 2021. Glyoxal as a potential source of highly viscous aerosol particles. *ACS Earth Space Chem* 5, 3324–3337. <https://doi.org/10.1021/acsearthspacechem.1c00245>.
- Qi, X., Zhu, S., Zhu, C., Hu, J., Lou, S., Xu, L., Dong, J., Cheng, P., 2020. Smog chamber study of the effects of  $\text{NO}_x$  and  $\text{NH}_3$  on the formation of secondary organic aerosols and optical properties from photo-oxidation of toluene. *Sci. Total Environ.* 727, 138632 <https://doi.org/10.1016/j.scitotenv.2020.138632>.
- Romonosky, D.E., Laskin, A., Laskin, J., Nizkorodov, S.A., 2015. High-resolution mass spectrometry and molecular characterization of aqueous photochemistry products of common types of secondary organic aerosols. *J. Phys. Chem. A* 119, 2594–2606. <https://doi.org/10.1021/jp509476r>.
- Sarrfrazadeh, M., Wildt, J., Pullinen, I., Springer, M., Kleist, E., Tillmann, R., Schmitt, S. H., Wu, C., Mentel, T.F., Zhao, D., Hastie, D.R., Kiendler-Scharr, A., 2016. Impact of  $\text{NO}_x$  and OH on secondary organic aerosol formation from  $\beta$ -pinene photooxidation. *Atmos. Chem. Phys.* 16, 11237–11248. <https://doi.org/10.5194/acp-16-11237-2016>.
- Shen, R.R., Liu, Z.R., Liu, Y.S., Wang, L.L., Li, D., Wang, Y.S., Wang, G.A., Bai, Y., Li, X.R., 2018. Typical polar organic aerosol tracers in  $\text{PM}_{2.5}$  over the North China Plain: spatial distribution, seasonal variations, contribution and sources. *Chemosphere* 209, 758–766. <https://doi.org/10.1016/j.chemosphere.2018.06.133>.
- Smith, J.D., Kinney, H., Anastasio, C., 2016. Phenolic carbonyls undergo rapid aqueous photodegradation to form low-volatility, light-absorbing products. *Atmos. Environ.* 126, 36–44. <https://doi.org/10.1016/j.atmosenv.2015.11.035>.
- Smith, N.R., Montoya-Aguilera, J., Dabdub, D., Nizkorodov, S.A., 2021. Effect of humidity on the reactive uptake of ammonia and dimethylamine by nitrogen-containing secondary organic aerosol. *Atmosphere* 12, 1502. <https://doi.org/10.3390/atmos12111502>.
- Volkamer, R., Ziemann, P.J., Molina, M.J., 2009. Secondary organic aerosol formation from acetylene ( $\text{C}_2\text{H}_2$ ): seed effect on SOA yields due to organic photochemistry in the aerosol aqueous phase. *Atmos. Chem. Phys.* 9, 1907–1928. <https://doi.org/10.5194/acp-9-1907-2009>.
- Wang, G., Zhang, R., Gomez, M.E., Yang, L., Levy Zamora, M., Hu, M., Lin, Y., Peng, J., Guo, S., Meng, J., Li, J., Cheng, C., Hu, T., Ren, Y., Wang, Y., Gao, J., Cao, J., An, Z., Zhou, W., Li, G., Wang, J., Tian, P., Marrero-Ortiz, W., Secrest, J., Du, Z., Zheng, J., Shang, D., Zeng, L., Shao, M., Wang, W., Huang, Y., Wang, Y., Zhu, Y., Li, Y., Hu, J., Pan, B., Cai, L., Cheng, Y., Ji, Y., Zhang, F., Rosenfeld, D., Liss, P.S., Duce, R.A., Kolb, C.E., Molina, M.J., 2016a. Persistent sulfate formation from London Fog to Chinese haze. *Proc. Natl. Acad. Sci. U.S.A.* 113, 13630–13635. <https://doi.org/10.1073/pnas.1616540113>.
- Wang, Y.J., Luo, H., Jia, L., Ge, S.S., 2016b. Effect of particle water on ozone and secondary organic aerosol formation from benzene- $\text{NO}_2$ -NaCl irradiations. *Atmos. Environ.* 140, 386–394. <https://doi.org/10.1016/j.atmosenv.2016.06.022>.
- White, J., S., Jamie M., L., Angove E., D., 2014. Chemical characterisation of semi-volatile and aerosol compounds from the photooxidation of toluene and  $\text{NO}_x$ . *Atmos. Environ.* 83, 237–244. <https://doi.org/10.1016/j.atmosenv.2013.11.023>.
- Wildt, J., Mentel, T.F., Kiendler-Scharr, A., Hoffmann, T., Andres, S., Ehn, M., Kleist, E., Müsgen, P., Rohrer, F., Rudich, Y., Springer, M., Tillmann, R., Wahner, A., 2014. Suppression of new particle formation from monoterpene oxidation by  $\text{NO}_x$ . *Atmos. Chem. Phys.* 14, 2789–2804. <https://doi.org/10.5194/acp-14-2789-2014>.
- Wu, C., Zhu, B., Liang, W., Ruan, T., Jiang, G., 2022. Molecular characterization of nitrogen-containing organic compounds in fractionated atmospheric humic-like substances (HULIS) and its relationship with optical properties. *Sci. Total Environ.* 832, 155043 <https://doi.org/10.1016/j.scitotenv.2022.155043>.
- Xie, M., Chen, X., Hays, M.D., Lewandowski, M., Offenberg, J., Kleindienst, T.E., Holder, A.L., 2017. Light absorption of secondary organic aerosol: composition and contribution of nitroaromatic compounds. *Environ. Sci. Technol.* 51, 11607–11616. <https://doi.org/10.1021/acs.est.7b03263>.
- Xu, L., Guo, H., Boyd, C.M., Klein, M., Bougiatioti, A., Cerully, K.M., Hite, J.R., Isaacman-Wertz, G., Kreisberg, N.M., Knote, C., Olson, K., Koss, A., Goldstein, A.H., Hering, S.V., de Gouw, J., Baumann, K., Lee, S.H., Nenes, A., Weber, R.J., Ng, N.L., 2015. Effects of anthropogenic emissions on aerosol formation from isoprene and monoterpenes in the southeastern United States. *Proc. Natl. Acad. Sci. U.S.A.* 112, 37–42. <https://doi.org/10.1073/pnas.1417609112>.
- Zhang, J., An, J., Qu, Y., Liu, X., Chen, Y., 2019a. Impacts of potential HONO sources on the concentrations of oxidants and secondary organic aerosols in the Beijing-Tianjin-Hebei region of China. *Sci. Total Environ.* 647, 836–852. <https://doi.org/10.1016/j.scitotenv.2018.08.030>.
- Zhang, Q., Xu, Y., Jia, L., 2019b. Secondary organic aerosol formation from OH-initiated oxidation of *m*-xylene: effects of relative humidity on yield and chemical composition. *Atmos. Chem. Phys.* 19, 15007–15021. <https://doi.org/10.5194/acp-19-15007-2019>.
- Zhang, R., Wang, G., Guo, S., Zamora, M.L., Ying, Q., Lin, Y., Wang, W., Hu, M., Wang, Y., 2015. Formation of urban fine particulate matter. *Chem. Rev.* 115, 3803–3855. <https://doi.org/10.1021/acs.chemrev.5b00067>.
- Zhang, W.Y., Wang, W.G., Li, J.L., Peng, C., Li, K., Zhou, L., Shi, B., Chen, Y., Liu, M.Y., Ge, M.F., 2020. Effects of  $\text{SO}_2$  on optical properties of secondary organic aerosol generated from photooxidation of toluene under different relative humidity conditions. *Atmos. Chem. Phys.* 20, 4477–4492. <https://doi.org/10.5194/acp-20-4477-2020>.
- Zhang, Y.Z., Forrister, H., Liu, J.M., Dibb, J., Anderson, B., Schwarz, J.P., Perring, A.E., Jimenez, J.L., Campuzano-Jost, P., Wang, Y.H., Nenes, A., Weber, R.J., 2017. Top-of-atmosphere radiative forcing affected by brown carbon in the upper troposphere. *Nat. Geosci.* 10, 486–489. <https://doi.org/10.1038/Ngeo2960>.
- Zhao, D., Schmitt, S.H., Wang, M., Acir, I.H., Tillmann, R., Tan, Z., Novelli, A., Fuchs, H., Pullinen, I., Wegener, R., Rohrer, F., Wildt, J., Kiendler-Scharr, A., Wahner, A., Mentel, T.F., 2018. Effects of  $\text{NO}_x$  and  $\text{SO}_2$  on the secondary organic aerosol formation from photooxidation of  $\alpha$ -pinene and limonene. *Atmos. Chem. Phys.* 18, 1611–1628. <https://doi.org/10.5194/acp-18-1611-2018>.
- Zou, Y., Deng, X.J., Zhu, D., Gong, D.C., Wang, H., Li, F., Tan, H.B., Deng, T., Mai, B.R., Liu, X.T., Wang, B.G., 2015. Characteristics of 1 year of observational data of VOCs,  $\text{NO}_x$  and  $\text{O}_3$  at a suburban site in Guangzhou, China. *Atmos. Chem. Phys.* 15, 6625–6636. <https://doi.org/10.5194/acp-15-6625-2015>.